

The Concentration of Potentially Toxic Elements in *Shrimp (Caridea)* Fillet from Iran: A Systematic Review and Potential Dietary Health Risk Assessment

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Abstract

Background: Pollution by potentially toxic elements (PTEs) have adverse effects on aquatic ecosystems. They may accumulate in sediments, plants, and aquatic organisms, disrupting the delicate balance of the ecosystem.

Methods: This study attempted to estimate the concentration of PTEs in shrimp tissue and estimate potential health risks in consumers among international databases, including Scopus, PubMed, Embase, and Web of Science, from January 2000 to March 20 2023. Furthermore, non-carcinogenic and carcinogenic risks for Iranian adult and child consumers were calculated using the Target hazard quotient (THQ) and Cancer Risk (CR).

Results: The rank order of PTEs due to THQ in adults was Methyl Hg (0.27) > Cd (0.13) > Ni (0.07) > Inorganic As (0.04) > Cu (0.03) > Pb (0.02), and for children Methyl Hg (1.28) > Cd (0.62) > Ni (0.35) > Inorganic As (0.20) > Cu (0.15) > Pb (0.11). TTHQ for the adults and children was 0.58 and 2.73, respectively. Besides, the mean CR in adults and children regarding inorganic As was 2.00E-5 and 7.98E-6, respectively. On the other hand, the total non-carcinogenic risk showed that children could be at considerable risk.

Conclusion: The carcinogenic risk of inorganic As was acceptable (CR < 1E-4) for adults and children. As a result, health and environmental authorities in Iran and other nearby regions must continually monitor the levels of PTEs in shrimp muscle tissue and implement policies to reduce the PTE pollution of the Persian Gulf and the Sea of Oman.

Keywords: Carcinogens, Cadmium, Shrimp, Heavy metals, Risk assessment

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Introduction

Water pollution due to Persistent Toxic Elements (PTEs) is a critical issue plaguing our global water systems, posing significant threats to environmental integrity and human health. PTEs such as lead (Pb), methyl mercury (Hg), cadmium (Cd), chromium (Cr), arsenic (As), and others are naturally occurring elements with high atomic weights, which can be toxic even at low concentrations (1,2). PTEs pollution of aquatic bodies may come from various sources, including mining, processing metal, and manufacturing, causing PTEs pollution. PTEs may enter water sources when industrial waste and wastewater

are improperly discharged (1). Moreover, agricultural practices that use fertilizers, pesticides, and livestock waste can introduce PTEs into water systems(1).

The PTEs pollution can have adverse effects on aquatic ecosystems. They may accumulate in sediments, plants, and aquatic species, disrupting the ecosystem's delicate balance (3). The reduced biodiversity may result from PTEs affecting the development and reproduction of aquatic plants and animals (1). Additionally, they may impact an organism's physiology and behavior, harming its immune system's capacity for reproduction (4). A notable historical example is the Minamata Bay



incident in Japan, where industrial discharge led to severe mercury contamination, significantly impacting aquatic life. Methylmercury, a type of mercury found in water, accumulates in fish, causing deformities, neurological harm, and reproductive problems in various species, including fish and shellfish. This pollution affected higher predators, such as birds and mammals, leading to compromised reproduction and population declines, notably affecting species like the endangered black-footed albatross. Furthermore, PTEs can potentially enter the food chain, posing risks to human health. When aquatic organisms, such as fish and shellfish, are contaminated with PTEs, consuming these contaminated organisms can lead to human exposure. Chronic exposure to PTEs through contaminated water or the consumption of contaminated food can lead to various health issues. PTEs can accumulate in human tissues and organs, leading to organ damage, neurological disorders, developmental abnormalities, and even cancer.

In Africa, as in other regions of the World, the continuous release of contaminants, including PTEs, PAHs, PCBs, plastics, pharmaceuticals, and other waste substances, into the marine ecosystem poses significant challenges for management solutions. Due to the stable, non-biodegradable, and persistent nature of these pollutants, which can accumulate in organs and tissues over time, human exposure to PTEs should be a significant concern. Reported health risks associated with human metal toxicity include kidney and skeletal damage, endocrine disruption, neurological disorders, cardiovascular dysfunction, and various cancers (5). Despite increasing evidence of PTEs pollution in the coastal waters of developing countries, there is a need for more data on PTEs levels in fish, daily intake levels, and the potential health risks to humans (6,7).

Based on an article published by the Food and Agriculture Organization (FAO), the production of various shrimp species worldwide exceeded 7.7 million tons in 2020 (8). In 2021, the global shrimp market reached USD 53.91 billion. Experts anticipate it will expand at a compound annual growth rate (CAGR) of 5.3% from 2022 to 2030. Economic progress, population growth, and cultural preferences for seafood have all impacted the trajectory of shrimp consumption in the Persian Gulf nations. These countries, including Iran, Saudi Arabia, Kuwait, the United Arab Emirates, and Qatar, have witnessed an increase in shrimp consumption as part of their diverse culinary traditions and growing demand for seafood products. Different species of shrimp have different benefits for consumption. White leg shrimp (*Litopenaeus vannamei*), Black tiger shrimp (*Penaeus monodon*), and Pink shrimp (*Farfantepenaeus duorarum*) are some different types of shrimp (9). Consuming shrimp offers several benefits due to its nutritional profile. It has few calories, includes essential vitamins and minerals, such as

selenium and omega-3 fatty acids, and is a rich source of high-quality protein (10,11). Additional health advantages of shrimp include enhanced heart health, enhanced cognitive function, and antioxidant properties (12).

Various studies and research have documented PTEs in fish and marine organisms in the Persian Gulf (13). These studies investigated the levels of PTEs in different fish species and other marine organisms collected from the Persian Gulf region. The concentrations of lead (Pb) and arsenic (As) in the muscles of shrimp from the Persian Gulf were found to exceed the guidelines set by the Food and Agriculture Organization/World Health Organization (FAO/WHO) (14). The mean concentrations of zinc and copper were lower than the international standards for the highest permissible limits (MPL)(15). A study by Al-Yamani et al. (2004) analyzed the concentrations of PTEs, including mercury (Hg), lead (Pb), cadmium (Cd), and copper (Cu), in different fish species from Kuwaiti waters, which are located in the northern part of the *Persian Gulf*. They found varying levels of these PTEs in the analyzed fish species (16). The concentrations of PTEs can vary depending on factors such as the species of marine organisms, their habitat, and the specific location within the *Persian Gulf*. In addition, the contamination of marine ecosystems and the accumulation of PTEs in organisms may be attributed to shipping, urbanization, and industrial activity along the Persian Gulf coastlines(16). The Persian Gulf stands out as one of the Middle East's most environmentally compromised marine ecosystems due to significant urbanization, industrial growth, and extensive development. The region faces heightened exposure and strain from oil exploration, extraction, and transportation (15). Additionally, the discharge of wastewater from nearby coastal cities, alongside petroleum refineries and fisheries operations, further exacerbates its environmental challenges. Researching PTEs in the Persian Gulf is crucial for understanding the extent of environmental pollution, identifying its sources and pathways, and assessing its potential impacts on the ecosystem and human health. Vigilant oversight and effective control of PTEs in this area are crucial for fostering sustainable growth, safeguarding marine biodiversity, and ensuring the welfare of communities that depend on the Gulf's resources. Several studies have been conducted on the concentration of PTEs in the Persian Gulf shrimps (17-21), but no systematic review study was found. This study aimed to determine the concentration of PTEs in the Persian Gulf shrimps and assess the associated health risks to consumers.

Materials and Methods

Searching strategy

The study was conducted according to the PRISMA protocol (22). The search was conducted by two authors (M. SA and Y. FA) in the international databases such

as Scopus, PubMed, Embase, and Web of Science from January 2000 to March 20, 2023. Keywords used for the search of papers were «Potential toxic element» OR «Metals» OR «Heavy metals» OR «Toxic element» OR «Trace metals,» AND «SHRIMP,» OR «Caridean Shrimp» OR «*Penaeus Vannamei*» OR «*Penaeus Indicus*» OR «*Penaeus Semisulcatus*» OR Caridea» OR «Marine food» and «Persian Gulf». The other articles were downloaded, and their entire texts were evaluated; only those that met our requirements were included in the data extraction stage. The articles' references were verified to recover any missing documents. The final correspondence comment resolved a disagreement among authors regarding the selection or removal of papers (23-25).

Inclusion/exclusion criteria and data extraction

The inclusion criteria were publication within the specified time frame, availability of complete texts in English, analysis of PTEs in shrimp from the Persian Gulf, use of a reliable detection technique, and provision of the range, average, or concentration of PTEs in shrimp fillets. Books, book chapters, review studies, correspondence, theses, conferences, and letters to editors were excluded. Statistics on the concentration of PTEs (inorganic As, methyl-Hg, Cd, Ni, Pb, and Cu), including average, standard deviation, lowest and highest values, the species of shrimp, and the detection method, were extracted from papers. Some studies reported total arsenic (As) levels in shrimp. Hence, the total As was converted to inorganic As using a 5% coefficient (26-28).

Health Risk Assessment

The daily intake from the ingestion of shrimp content of PTEs was calculated using equations below:

$$CDI = \frac{C \times IR \times ED \times EF}{BW \times AT} \quad (1)$$

Where *CDI* is chronic daily intake; *C* is the concentration of PTEs in shrimp; *IR* is ingestion rate; *EF* is exposure frequency (350 d/y); *ED* is exposure duration (children and adults are 6 and 70 y, respectively); and *BW* is body weight (children and adults are 20 and 70 kg, respectively). The average lifetime non-carcinogenic risk for children and adults is 2190 days, and the average lifetime carcinogenic risk for both children and adults is 25550 days. The average per capita consumption of marine foods in Iran is 13.4 kg/y (29). On average, 15% of marine food consumption is dedicated to shrimp (12,30). Hence, the ingestion rate of shrimp in Iran can be as high as 2.01 kg/year. The RfD for methyl Hg, inorganic As, Ni, Cd, and Cu are 0.0001, 0.0003, 0.11, 0.001, and 0.04 mg/kg-d, respectively (31). The tolerable daily intake for Pb is 0.0036 mg/kg-d (32).

The non-carcinogenic risk was calculated by the target hazard quotient equation (32):

$$THQ = \frac{CDI}{RfD \text{ or TDI}} \quad (2)$$

Where *THQ* is the target hazard quotient.

When $THQ < 1$, the exposed populations are at an acceptable non-carcinogenic risk.

The carcinogenic risk of inorganic As was calculated using the following equation (32):

$$CR = CDI \times SF \quad (3)$$

Where *CR* is Carcinogenic risk, *CDI* is chronic daily intake, and *SF* is slope factor.

SF for inorganic As equals 1.5 (mg/kg-d)⁻¹. When $CR < 1.00E-4$ value, the exposed population is an acceptable carcinogenic risk (32).

Results

The concentration of PTEs

Of the 754 articles obtained, 673 were excluded in the initial review. According to titles and abstracts, 81 articles were identified as potentially suitable. Sixty-five articles were excluded due to the detection of other marine foods and elements; no original data were reported (review, book, chapter, letters, and conferences). After being evaluated for eligibility, 16 papers with 54 data reports published between January 2000 and March 20, 2023, were included in our study (Figure 1 and Table 1) for meta-analysis and carcinogenic risk assessment.

The highest concentration of inorganic As (1.55 mg/kg) and Cd (14.00 mg/kg) were observed in the study by Soltani et al. (Figure 2). The highest concentration of methyl-Hg (3.64 mg/kg) was observed in the study by Rahmanpour et al. (Figure 3). The highest concentration of Pb (4834 mg/kg) was observed in the study by Monikh et al. (Figure 3).

The highest concentrations of Ni (89.32 mg/kg) and Cu (97.79 mg/kg) were observed in the studies by Rahmanpour et al. and Pourang et al., respectively (Figure 4). The rank order of PTEs based on the percentage of THQ higher than 1 value was Cd (4.60%) > Methyl Hg (0%) ~ inorganic As (0%) ~ Pb (0%) ~ Ni (0%) ~ Cu (0%) for adults, and Cd (16.27%) > Ni (14.81%) > Methyl Hg (11.11%) > Inorganic As (6.60%) > Pb (0%) ~ Cu (0%) for children (Table 2).

Risk assessment

Based on the data presented in Table 3, the rank order of PTEs based on the THQ in adults was Methyl Hg (0.27) > Cd (0.13) > Ni (0.07) > Inorganic As (0.04) > Cu (0.03) > Pb (0.02), and for children, Methyl Hg (1.28) > Cd (0.62) > Ni (0.35) > Inorganic As (0.20) > Cu (0.15) > Pb (0.11). TTHQ for adults and children was 0.58 and 2.73, respectively (Figure 5). Also, the percentage of CR higher than 1E-6 for the adults and children was equal to 0.00% and 0.00%, respectively (Table 2). The mean CR in adults

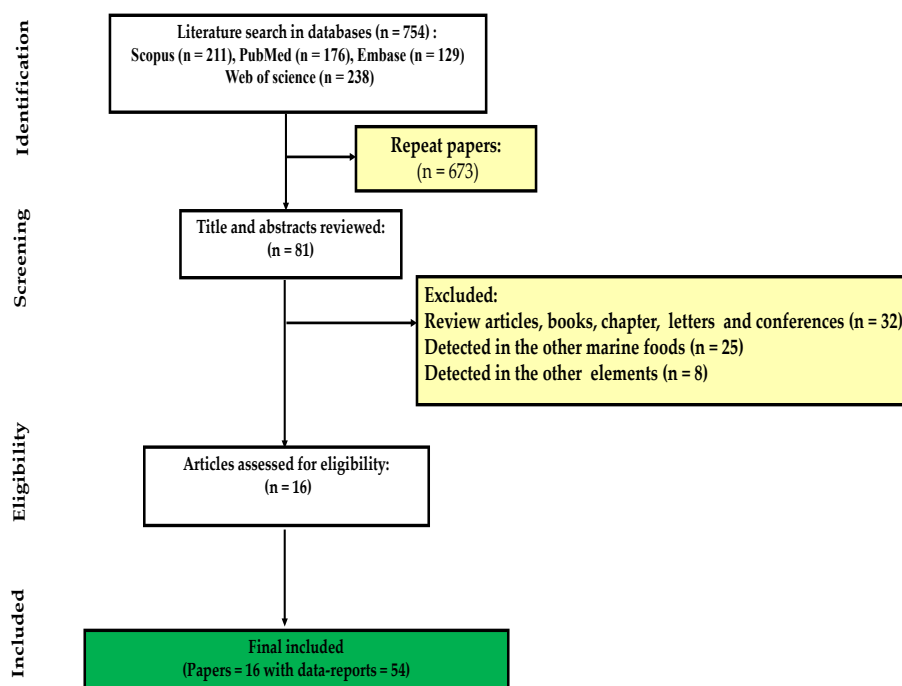


Figure 1. Process of selecting papers based on the PRISMA

and children due to inorganic As was 2.00×10^{-5} and 7.98×10^{-6} , respectively (Figure 6).

Discussion

The highest concentration of inorganic As (1.554 mg/kg) and Cd (14.00 mg/kg) was observed in the study by Soltani *et al.* (Figure 2). These findings raise concerns about the discharge of industrial effluents caused by the activities of various industries in the region. Nevertheless, in all studies, the mean concentrations of arsenic (As) were below the Maximum Permissible Limit (MPL) of 2 mg/kg, as provided by Food Standards Australia and New Zealand (48). However, in 79% of studies, the mean concentrations of Cd exceeded the UN permissible amount by the European Commission (0.05 mg/kg) (49). These results can be attributed to human activities (e.g., oil refineries, fishing) in coastal areas (50). The major source of As in Khor Musa may be the discharge of industrial effluents from the operations of numerous enterprises in the area, particularly petrochemical complexes that employ arsenic catalysts (51). Nyarko *et al.* (2023) reported a higher concentration of mean As for fish species (0.8 to 8.5 mg/kg) and 0.002 mg/kg to 0.027 mg/kg of Cd in the fish species in the Gulf of Guinea that were in the MPL of the EU. The high As levels in this study may be due to the use of herbicides in mining and agricultural activities (52). However, Olmedo *et al.* (2013) found As and Cd levels in fresh shrimp (*Parapenaeus longirostris*) at 0.739 and 0.029 mg/kg ww, respectively (53). PTEs are considered an indicator of pollution because their exposure to water bodies leads to a decrease in the quality of water ecosystems (54). The bioaccumulation of metals in fish

and shrimp may vary across species and is influenced by dietary preferences, ecological requirements, sex, age, size, physical and chemical properties of water, seasonal fluctuations, and the concentration of metals in water and sediments. Among PTEs, arsenic (As), cadmium (Cd), mercury (Hg), lead (Pb), and copper (Cu) have the highest potential for accumulation in fish tissue (55,56). Arsenic tends to accumulate in sediments, so the creatures of the deep sea are exposed to larger amounts of this pollutant due to their proximity to sediments (57). Exposure to very small amounts of arsenic may act as an endocrine disruptor. In contrast, long-term exposure to inorganic arsenic can lead to problems in the digestive, respiratory, skin, liver, vascular, and nervous systems (58).

Cd is a dangerous non-essential metal (59) that can cause chronic toxicity to various organisms, even at low concentrations (54,57). Certain environmental factors, including temperature, calcium concentration, and salinity, can influence the toxicity of this metal (59). Despite its low background concentration in the environment, exposure to cadmium (Cd) from smelting, electroplating, or even fertilizer consumption can cause kidney failure, bone softening, itai-itai disease, anemia, olfactory nerve damage, and even prostate cancer (60).

This can be caused by the accumulation of this metal in the sediments of Khuzestan province, resulting from petrochemical activities such as chlor-alkali production in the region. The Arvand River, the main source of freshwater entering the Persian Gulf, can dilute salty seawater and contribute to the formation of sediments. Therefore, mean Hg concentrations in *Metapenaeus affinis* species were higher than the MPL (0.05 mg/kg) set by

Table 1. Main characteristics included in the present study (mg/kg-ww)

Species	Sample size	Inorganic As				Cd				Methyl-Hg				Pb				Ni				Cu				Method of detection	Ref.
		Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max		
shrimps (<i>Penaeus merguensis</i>)	60	0.009	0.002			0.194	0.044			0.063	0.021			0.325	0.143											FIAS 4100, Perkin-Elmer	(33)
lobsters (<i>Panulirus homarus</i>)	60	0.005	0.003			0.163	0.077			0.053	0.012			0.641	0.084											FIAS 4100, Perkin-Elmer	(33)
<i>Fenneropenaeus merguensis</i>	21	0.001	0.001			0.037	0.001			0.007	0.000			0.087	0.003							1.260	0.200			Perkin-Elmer, model 4100 ZL atomic absorption spectrophotometer	(34)
<i>Penaeus merguensis</i>	20					0.072	0.126							0.121	0.139							19.385	7.562			ICP-AES	(35)
<i>Penaeus merguensis</i>	NM					0.381	0.207							0.129	0.179							97.798	39.042			ICP-AES	(35)
<i>Penaeus merguensis</i>	NM					0.016	0.006							0.076	0.116							17.861	11.527			ICP-AES	(35)
<i>Metapenaeus affinis</i>	20					0.098	0.231							0.231	0.179							28.198	48.838			ICP-AES	(35)
<i>Metapenaeus affinis</i>	NM					0.171	0.314							0.446	1.084							90.526	73.466			ICP-AES	(35)
<i>Metapenaeus affinis</i>	NM					0.023	0.013							0.245	0.361							17.369	9.868			ICP-AES	(35)
<i>Penaeus semisulcatus</i>	40					0.736	0.538	0.100	2.760													20.295	10.593	5.170	50.830	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
White Shrimp (<i>Litopenaeus vannamei</i>)	20													0.533	0.067			0.061	0.036			4.140	0.650			AAS (Perkin-Elmer Model Analyste 300)	(19)
White Shrimp (<i>Litopenaeus vannamei</i>)	20													0.002	0.025			0.082	0.021			0.950	0.240			AAS (Perkin-Elmer Model Analyste 300)	(19)
White Shrimp (<i>Litopenaeus vannamei</i>)	20													0.124	0.021			0.065	0.002			0.190	0.100			AAS (Perkin-Elmer Model Analyste 300)	(19)
White Shrimp (<i>Litopenaeus vannamei</i>)	20													0.069	0.029			0.080	0.002			0.060	0.010			AAS (Perkin-Elmer Model Analyste 300)	(19)
<i>Penaeus semisulcatus</i>	180													0.231	0.019	0.006	0.981	1.432	0.021	0.347	2.205					HR-ICP-MS	(37)
<i>Metapenaeus affinis</i>	27					0.021	0.000							0.004				0.025	0.000			0.810				atomic absorption	(37)
<i>Metapenaeus affinis</i>	30	0.022	0.006			1.430	0.540			1.320	0.210			3.050	0.700			51.920	7.830							ICP-AES	(38)
<i>Metapenaeus affinis</i>	30	0.007	0.002			0.110	0.020			0.210	0.010			0.760	0.310			52.150	13.470							ICP-AES	(38)

Table 1. Continued.

Species	Sample size	Inorganic As				Cd				Methyl-Hg				Pb				Ni				Cu				Method of detection	Ref.
		Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max		
Metapenaeus affinis	30	0.053	0.017			4.920	1.050			0.620	0.160			3.940	0.630			89.320	18.610							ICP-AES	(38)
Metapenaeus affinis	30	0.038	0.008			5.150	2.280			3.640	0.420			3.160	0.850			53.750	9.230							ICP-AES	(38)
Metapenaeus affinis	50					0.557	0.046							2.463	0.263			7.235	0.918			23.630	3.390			AAS	(39)
Metapenaeus affinis	50					0.647	0.038							3.629	0.470			8.875	1.560			25.610	3.640			AAS	(39)
Metapenaeus affinis	50					1.184	0.143							4.834	0.937			13.287	2.144			47.590	6.620			AAS	(39)
Metapenaeus affinis	50					0.053	0.006							1.428	0.328			6.764	0.998			16.080	2.890			AAS	(39)
Metapenaeus affinis	50					0.067	0.008							1.386	0.416			4.255	0.668			18.110	3.550			AAS	(39)
Metapenaeus affinis	50					0.032	0.063							2.898	0.544			5.200	0.363			31.750	4.620			AAS	(39)
-	NM					9.850	0.910			0.003	0.370			2.800	0.380							0.480	0.050			AA240	(40)
-	NM					10.000	0.750			0.003	0.290			2.200	0.250							0.440	0.300			AA241	(40)
-	NM					7.980	0.570			0.003	0.480			2.160	0.230							0.470	0.030			AA242	(40)
Ghost shrimp, U. pseuduchelata	7					0.720	1.060							1.500	1.200											GFAAS	(41)
Ghost shrimp, U. carinicauda	29					0.130	0.030							0.630	0.330											GFAAS	(41)
Ghost shrimp, M. indica	16					0.280	0.048							0.090	0.031											GFAAS	(41)
Ghost shrimp, P. bouvieri	23					0.210	0.020							1.026	0.420											GFAAS	(41)
Ghost shrimp, C. typa	4					0.042	0.010							0.020	0.007											GFAAS	(41)
Ghost shrimp, N. calmani	25					0.407	0.050							0.400	0.080											GFAAS	(41)
Penaeus semisulcatus	60	0.013	0.002							0.082	0.019															Perkin Elmer Model 4100 atomic absorption spectrometer	(42)
Panulirus homarus	60	0.009	0.002			0.251	0.070			0.049	0.009			0.629	0.169											Perkin Elmer Model 4100 atomic absorption spectrometer	(43)
Metapenaeus affinis	95	1.554	0.281	1.355	1.754	14.000				0.053	0.002	0.048	0.055	0.326	0.103	0.252	0.399	0.124	0.059	0.082	0.166	20.050	3.180	17.800	22.300	ICP-MS, Agilent 7700, USA	(44)
Penaeus semisulcatus	62	0.656	0.150	0.550	0.762	14.000				0.160	0.069	0.111	0.208	0.443	0.019	0.420	0.462	0.132	0.042	0.103	0.162	34.150	7.000	29.200	39.100	ICP-MS, Agilent 7700, USA	(44)

Table 1. Continued.

Species	Sample size	Inorganic As				Cd				Methyl-Hg				Pb				Ni				Cu				Method of detection	Ref.	
		Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max	Mean	sd	min	max			
Metapenaeus affinis	151	0.143	0.013	0.122	0.163	0.080					0.032	0.008	0.025	0.042	1.604	1.145	0.389	2.663	0.225	0.082	0.132	0.277	16.330	0.580	16.000	17.000	ICP-MS & FIMS	(45)
Penaeus semisulcatus	103	0.114	0.017	0.094	0.126	0.080					0.109	0.063	0.038	0.158	1.340	0.756	0.491	1.945	0.168	0.038	0.126	0.202	25.670	3.060	23.000	29.000	ICP-MS & FIMS	(45)
Metapenaeus affinis	80					0.027	0.001							0.032	0.001							10.000	0.060			ICP- MS, Elan 9000, Perkin Elmer, USA	(46)	
Metapenaeus affinis	NM					0.021	0.001							0.025	0.001							3.000	0.030			ICP- MS, Elan 9000, Perkin Elmer, USA	(46)	
Metapenaeus affinis	NM					0.021	0.002							0.017	0.001							5.500	0.090			ICP- MS, Elan 9000, Perkin Elmer, USA	(46)	
Metapenaeus affinis	NM													0.002	0.001							1.100	0.040			ICP- MS, Elan 9000, Perkin Elmer, USA	(46)	
Penaeus merguensis	180													0.021	0.006			1.283	0.029							HR-ICP-MS	(47)	
Penaeus semisulcatus	40					0.573	0.339	0.007	1.020														11.224	2.878	7.430	21.020	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
Penaeus semisulcatus	40					0.191	0.274	0.001	1.210														3.384	0.881	1.290	5.460	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
Penaeus merguensis	40					0.573	0.422	0.140	2.790														22.232	13.237	6.460	83.480	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
Penaeus merguensis	40					0.418	0.299	0.010	0.890														13.354	5.102	4.060	28.430	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
Penaeus merguensis	40					0.076	0.051	0.010	0.180														4.873	3.671	1.420	23.520	ICP-AES; Philips/Pye Unicam Model PU 7000 & ICP-MS; VG PlasmaQuad 3-VG Elemental	(36)
Penaeus semisulcatus	60	0.013	0.002							0.082	0.017															Perkin Elmer Model 4100 atomic absorption spectrometer	(42)	

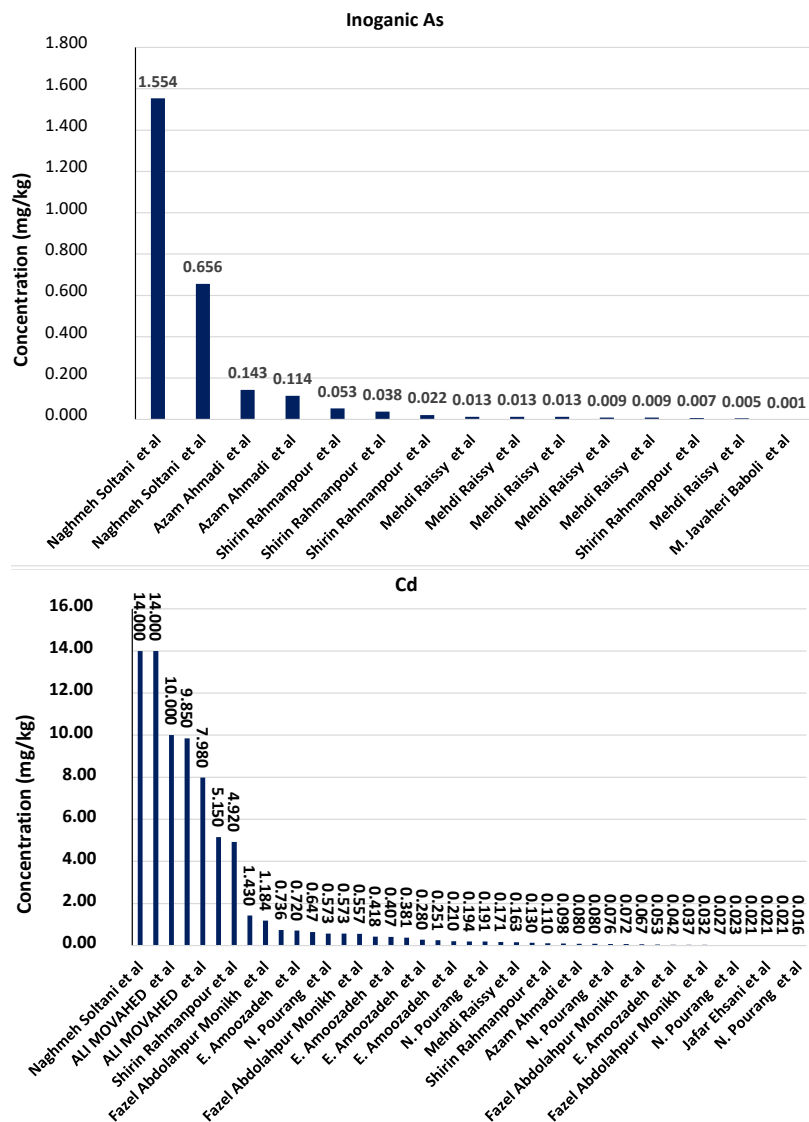


Figure 2. Concentrations of inorganic As and Cd in shrimp in the Persian Gulf

the EU. Compared to all studies, about 67% exceeded the MPL of the EU (61). Mercury is one of the rare elements, and 75-90% of the total mercury in fish is found in the dangerous organic form of methylmercury (58). Mercury in large quantities in the sediments of Khuzestan province, which serve as a trap for PTEs, is thought to cause the biological accumulation of mercury in aquatic ecosystems. Regarding differences in environmental and biological factors, PTEs in various ecosystems exhibit distinct accumulation patterns, even when similar species (62). Comparing the Kalogeropoulos et al. study, Hg levels (< 0.008 – 0.18 mg/kg) in raw Mediterranean finfish in Greece appear to be noticeably lower than those (63).

The highest concentration of Pb (4.83 mg/kg) was observed in the Abdolalpour et al. study (Figure 3), which was much higher than the MPLs set by the Codex Alimentarius Commission (WHO/FAO 2019) (0.300 mg/kg) (64). Approximately 36% of studies reported Pb concentrations within the WHO/FAO range. This result

is consistent with the results of research by Herliwati et al. (2022), which revealed substantial Pb contamination (3.150 mg/kg) in samples taken from the Barito River estuary in Indonesia, indicating the accumulation of heavy metals from the water of the area to the sediments and finally to the shrimp (65). Pb levels in muscles, exoskeleton, and gills of the prawn *Penaeus semisulcatus* in the study by El Gendy et al. (2015) were 2.33, 17.00, and 21.33 ppm, respectively (66). One of the reasons for the increase in Pb levels in seawater may be the provision of sailing and shipping services between Ports and Islands in the region (67). Lead is an unnecessary element that is very toxic and harmful, even in small amounts. It can lead to neurotoxicity, kidney toxicity, and other negative consequences, which can negatively affect human health if consumed in terms of accumulation in the tissues of aquatic organisms (58). It may harm human health if taken, due to the accumulation in the tissues of aquatic species. Pb may cause cognitive disabilities, renal damage,

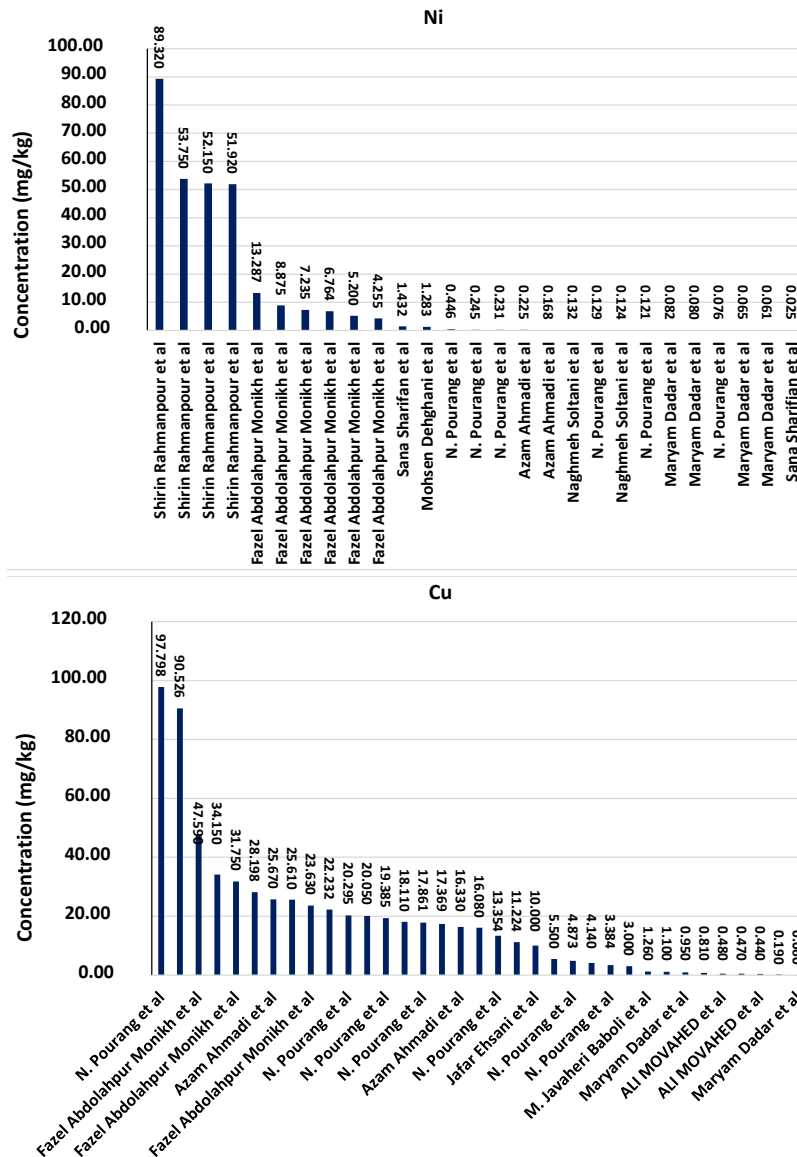


Figure 4. Concentrations of Ni and Cu in shrimp in the Persian Gulf

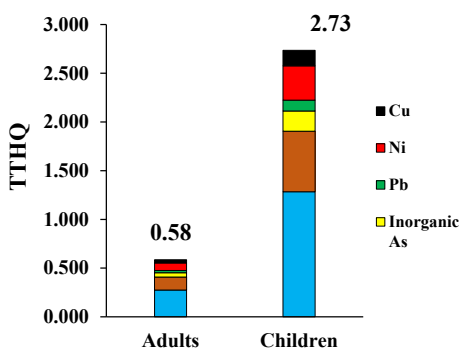


Figure 5. TTHQ in adults and children due to heavy metals in shrimp in the Persian Gulf

mg kg⁻¹ (70). Ezemonye et al. (2019) reported that Ni and Cu had mean values of 53.57 and 60.83 mg/kg in shrimp (*Macrobrachium macrobrachion*). Cu may have entered the water body through surface runoff and destruction of abandoned vessels near the coastlines adjacent to

petrochemical fields (71). These values were lower than shrimps present in the Persian Gulf.

Copper is a significant and essential metal in cell metabolism, growth, and survival. As a result, if relatively large amounts of this metal are detected in tissues, it is necessary (72). Excessive consumption of copper can lead to kidney and liver damage, vomiting, and gastrointestinal distress (58,70,73). Food is the main source of exposure of the general population to Ni (70). Although Ni is considered an essential heavy metal, it can be toxic to humans in large amounts. One of the primary sources of high amounts of nickel is the presence of petroleum substances, such as gasoline, engine oil, and diesel fuel, that leak from motorboats on the seashore. The effects of nickel on health include hair loss and worsening of eczema, as well as adverse effects on the kidneys, cardiovascular system, immune system, and reproductive system (54).

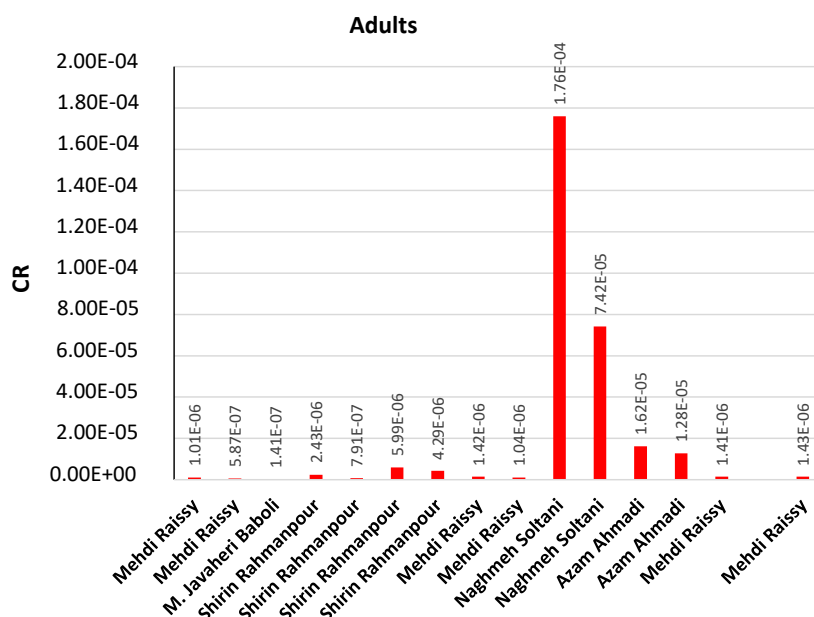


Figure 6. Mean cancer risks (CR) in adults and children due to inorganic As in shrimp of the Persian Gulf

Health risk assessment

Microbial and chemical contamination in food can have detrimental effects on consumer health (25). The non-carcinogenic and carcinogenic effects on humans can be due to their toxic nature, stable properties, and the accumulation of toxic metals due to their biological accumulation properties. As mentioned earlier, if the $THQ > 1$, it implies the non-carcinogenic health threats for the exposed population (74)

As can be seen, the non-cancer risk caused by cadmium is more evident in the studies, which can be in terms of relatively high concentration and the very high toxicity of cadmium. The percentage of $THQ > 1$ in children was higher than in adults, likely due to their lower body weight, which may lead them to take risks. These findings imply that children's and adults' long-term ingestion of mercury and cadmium from shrimp poses the most significant non-cancer health hazards because of their increased toxicity. Our findings align with those of Nyarko et al, who also reported elevated As and Cd levels in fish from the Gulf of Guinea. The estimated THQ of As for *P. notialis* (Southern pink shrimp) in both children and adolescents, as well as the case of Cd in children, exceeded 1 (52). However, Kalogeropoulos et al reported that THQ for Pb, Cd, and Hg in Pan-fried shrimp had no health risks to consumers (63). TTHQ for children was higher than 1, indicating a potential risk to children's health from consuming shrimp. Ezemonye et al. found a similar TTHQ (above 1) in their study for shrimp consumption (*Macrobrachium macrobrachion*) (71). However, the study by Ahmed et al showed TTHQ levels below the permissible level of 1 for shrimp (*Macrobrachium rosenbergii* and *Metapenaeus monoceros*) (75). This difference can be attributed to different concentrations of elements in shrimp and their

different per capita consumption. Usually, Cancer Risks (CR) higher than $1E-4$ are considered considerable (76). These results showed that the CR values of As were not considerable for adults and children. These results are based on the research by Ahmed et al which showed that the CR was between 10^{-7} and 10^{-5} for all shrimp samples (75). Although children are lighter than adults, adults still have a higher chance of developing cancer because children's contact time with the environment is estimated to last 70 years, making children's carcinogenesis risk significantly lower.

Conclusion

This study attempted to collect studies related to the concentration of PTEs in shrimp consumed in Iran and estimate the potential health risk. The concentration of PTEs in shrimp was observed to be different. In some studies, the concentrations exceeded the standard limits, while in others, they were below them. This difference in concentration is attributed to variations in shrimp genus and species, sampling location, and sampling season. According to the health risk assessment, Methyl Hg and Cd may pose a larger non-carcinogenic risk than other PTEs. On the other hand, the overall non-carcinogenic risk demonstrated that youngsters may be exposed to a high level of danger. Thus, the carcinogenic risk of inorganic As was acceptable ($CR < 1E-4$) in adults and children. In future studies, the difference in the concentration of PTEs in the Persian Gulf and Sea of Oman shrimps and other pollutants, such as antibiotics, pesticides, and PAHs in shrimp tissue, can be considered.

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Competing interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Ethical issues

All methods were carried out in accordance with relevant guidelines and regulations, and all participants have read and declared their consent in accordance with relevant guidelines. Written informed consent was obtained from the participants before data collection.

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